

ARTICLE

Freshwater Ecology

The benefits of merging passive and active tracking approaches: New insights into riverine migration by salmonid smolts

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Abstract

The process of smolting is a critical phase in the life cycle of anadromous salmonids, and it has been associated with substantial rates of mortality. Survival during freshwater and marine migration is known to have population-level effects; thus, an understanding of the patterns of mortality has the potential to yield important insights into population bottlenecks. Despite important advancements in tracking techniques, the specifics of mortality events in anadromous salmonids during their initial migration to sea remain somewhat elusive. Here, we develop a framework combining spatial and temporal detections of smolt riverine migration from two tracking techniques, which enable inferences to be made about mortality locations, causes, and rates. In this study, we demonstrate that during their initial riverine transitional phase, smolts were particularly vulnerable to predators. Specifically, avian predation appeared to be the main cause of mortality (42%), although piscine predation events were not trivial (14%). Our results suggested some direct and indirect tagging-induced mortality (e.g., through increased predation vulnerability), which highlights the importance of determining tagging mortality in a telemetry study to ensure adequate interpretation of migration success. Overall, by estimating migration loss and its variability, our study framework should help to guide management actions to mitigate the widespread population declines these species are currently facing.

KEYWORDS

acoustic tags, anadromous, Atlantic salmon, mortality, predation, river, sea trout, small-scale behavior, telemetry

INTRODUCTION

Atlantic salmon (*Salmo salar*) and sea trout (the anadromous form of the brown trout complex; *Salmo trutta*) are

salmonid species of biological, cultural, and economic importance, but these species have experienced dramatic declines in recent decades (ICES, 2011; ICES, 2018). Despite various attempts to mitigate anthropogenic

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stressors and mortality rates, Atlantic salmon and sea trout populations throughout most of geographic range of these fish are currently at, or near, record low abundance (ICES, 2017, 2018). The anadromous life-history tactic of both species results in migration between freshwater and marine habitats, which exposes individuals to multiple threats (e.g., migration barriers, diseases, pollution, overexploitation, parasites, climate change, and aquaculture impacts; Forseth et al., 2017; Parrish et al., 1998). Smolts (the seaward migrating stage of the life cycle) are particularly vulnerable during their first migration to sea because they are increasingly mobile and traverse a high-risk landscape that exposes them to predators (Ward et al., 2008; Ward & Hvidsten, 2011). However, the location, timing, and relative proportion of the total mortality attributable to Atlantic salmon and sea trout populations during their initial migration remain elusive (Chaput et al., 2018). This lack of an understanding of migration losses is highly challenging for practical management and for policy development.

Recent advances in the development of electronic tags that transmit information that enables tracking of animals across broad and fine spatial and temporal scales have made significant contributions to the understanding of salmonid behavior and migration (Drenner et al., 2012; Hussey et al., 2015). Our ability to track aquatic organisms often varies with the species studied and the habitat they inhabit. There are trade-offs among operating requirements, transmission quality, and tracking capabilities in the devices used to track animals; these become more complex in species that migrate between freshwater and marine systems (Leander et al., 2019). Currently, electronic tags are the leading technology to track fish with high temporal and spatial resolution in the wild (Hussey et al., 2015). Radio telemetry has been commonly used in rivers, since it allows the determination of fine-scale movements and mortality (Jepsen et al., 2019; Keefer et al., 2012; Wertheimer & Evans, 2005). However, radio telemetry cannot be used to track animals in saltwater or deep-water environments, which is a major limitation when studying an anadromous migration (Cooke et al., 2013).

When freshwater and marine migrations by anadromous salmonids are investigated, passive acoustic telemetry offers a good alternative (McMichael et al., 2010). The main issue with passive acoustic telemetry is that a tagged individual must move within the detection range of an acoustic receiver to be detected (Both et al., 2012; Bruneel et al., 2020). This issue can be overcome by deploying a sufficient number of receivers to achieve complete coverage of a given area. For example, when the direction of migration is known or constrained in certain areas, using an array of receivers in a grid system

in relatively enclosed study areas (e.g., lakes and estuaries) and using an array of receivers organized as a gate through which fish are likely to travel, have both been proved to be efficient methods to detect organisms (Heupel et al., 2006; Kraus et al., 2018; Roy et al., 2014). Despite being well suited for marine or lake environments, these solutions to increase the resolution of the data do not necessarily work well in river systems. Rivers are often shallow, sinuous, and frequently with fast-flowing water that results in a noisy environment that makes tag detection challenging; thus, increasing the number of receivers for better detection resolution is not necessarily feasible (Bergé et al., 2012). Combining approaches with different trade-offs (e.g., using PIT tagging or radio telemetry along with acoustic telemetry) may circumvent some of the challenges of tracking in riverine systems, but would require double-tagging individuals or running programs in parallel (Dainys et al., 2018; Furey et al., 2016; Jepsen et al., 2019; Schwinn et al., 2017).

As a result of some of these shortcomings from passive acoustic telemetry associated with rivers, we revisited the concept of active acoustic tracking in association with passive acoustic telemetry. Active acoustic tracking is not a new concept (e.g., Gauld et al., 2016; Halfyard et al., 2012). Active acoustic tracking has been considered to be a detection method with some advantages; it is likely to provide: (1) a high frequency of detections of animal positions, (2) position estimates that are not limited to areas that are in the range of fixed receiver stations, and (3) relatively precise animal positions (Brownscombe et al., 2019). However, because active tracking is labor-intensive and as it often restricts the sample size compared with a passive tracking approach (e.g., frequently only one animal can be tracked at a time and the duration over which animals can be tracked is usually limited), methods using fixed acoustic receivers have been favored in telemetry studies (Brownscombe et al., 2019; Fetterplace et al., 2016). Novel active tracking techniques have been recently developed, such as autonomous robotic technology or receivers attached to large, free-ranging animals to enable detection of individuals of the target species (Carlson, 2015; Ennasr et al., 2020), none of which are currently suitable for most rivers. Thus, to counter the limitations of existing passive telemetry in small- to moderate-sized riverine systems and enhance the collected information on smolts during their freshwater migration phase, we developed a novel framework that integrates fixed acoustic receivers and active tracking by systematic canoe transects. The purposes of our study were to: (1) test the relative efficiency of each tracking methodology (i.e., passive and active) separately and together; (2) use the combined active-passive approach to define and quantify the locations, causes,

and rates of mortality of salmonid smolts during their riverine migration phase; and (3) determine the fine-scale behavior related to migration (e.g., time of day traveling and residency).

MATERIAL AND METHODS

Study site and tagging procedures

The River Endrick (Scotland) is a major inflowing tributary to Loch Lomond, a medium-sized lake in the catchment through which fish must migrate to reach the sea (Figure 1). Wild Atlantic salmon and sea trout smolts (there is no contemporary stocking of these species there) were captured using a 1.2-m-diameter rotary screw trap in the middle reaches of the River Endrick, 12.33 km upstream of the river mouth ($56^{\circ}2'58''$ N, $004^{\circ}26'27''$ W) where it discharges into Loch Lomond. The rotary screw trap was in constant operation (both day and night) over the study period. Between 13 April and 20 April 2020, a total of 135 Atlantic salmon and 23 sea trout smolts were tagged with VEMCO V7-2L 69 kHz (VEMCO Ltd, Halifax,

Canada) tags. To minimize the tag burden, only smolts in excess of 130-mm fork length were tagged. Smolts were anesthetized using 0.1 g L^{-1} of tricaine methanesulfonate, measured for weight (in grams) and length (fork length, in millimeters). A minimum recovery period of 1 h was allowed before fish were released at the tagging site.

Passive acoustic tracking design

To assess smolt migration through the River Endrick, four 69-kHz acoustic receivers (VEMCO VR2W and VR2Tx VEMCO Ltd.) were deployed in the river between 24 March and 14 April 2020. The first receiver was located 800 m downstream from the trapping/release zone, and the last receiver was placed at the mouth of river, to detect smolts entering Loch Lomond (1, 5, 6, and 7; Figure 1). On 21 May 2020, three additional acoustic receivers (VR2Tx; VEMCO Ltd.) were deployed in the river when the environmental conditions in the river (i.e., the lack of precipitation) resulted in a total obstruction of the smolt migration due to an exposed sandbank at the junction of the river mouth and Loch Lomond.

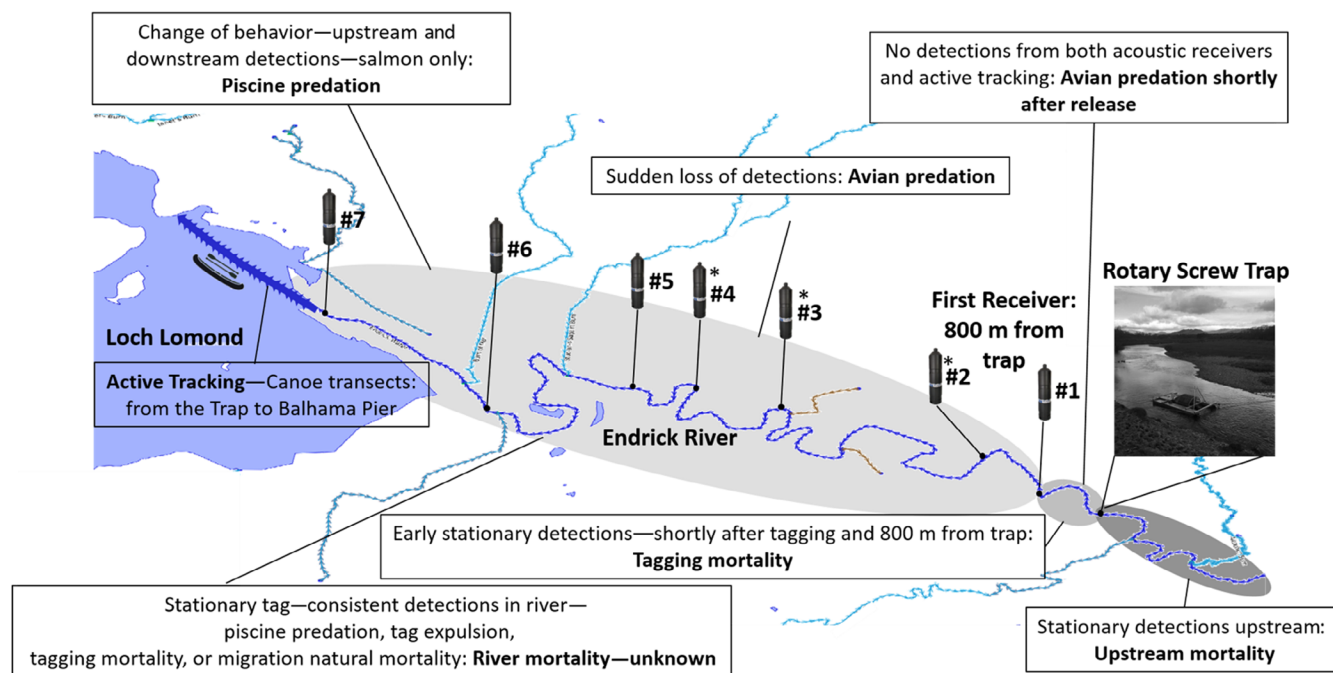


FIGURE 1 Schematic of the study area and the combined passive and active tracking methodology developed to monitor Atlantic salmon and sea trout smolt migration. The passive acoustic design includes four VR2W receivers initially installed between 24 March and 14 April 2020 (i.e., prior to fish tagging). Three additional receivers were installed on 21 May 2020, which are represented by an asterisk. The active tracking design consists of a canoe transect commencing ~200 m above the rotary screw trap (and the release site) down river to Balmaha Pier, beyond the mouth of the river, using a VR100 portable acoustic receiver. The River Endrick is represented in dark blue. The inflows to the River Endrick are represented in light blue, whereas the outflows are in yellow. Inferences drawn from the patterns of detections (or the lack of detections) from the combined active and passive acoustic receivers led to six plausible migration outcomes for each fish (beside the “unknown”). These inferences are presented in Table 1 and Appendix S1: Table S1

These three additional acoustic receivers (2–4) were deployed in the midsection of the river to expand detection coverage from the trapping zone to the mouth of the river (Figure 1). All receivers were attached to a steel bar welded onto a 20-kg weight, which was then anchored to the shore with a rope or chain (Appendix S1: Figure S1).

Active acoustic tracking design

The active tracking program consisted of a number of transects of 13.5 km, undertaken by canoe that began ~200 m upstream of the rotary screw trap, downstream into Loch Lomond (Balmaha Pier; 56°08'36" N, 004°54'79" W). Tags were detected during the passage downstream using a VR100 receiver (VEMCO Ltd.; normal filter setting), designed for manual tracking from small boats, coupled with an omnidirectional hydrophone, which was attached to the canoe. The operator could adjust the length of the cable to prevent the trailing hydrophone hitting the substrate. Canoe transects were completed daily, beginning 3 days after tagging commenced, for 7 days (i.e., 16 April to 23 April 2020) and every 2 weeks after that (i.e., 24 April to 6 June 2020), for an overall total of 10 active tracking surveys. The last canoe transect, completed on 6 June 2020, was started further upstream near a waterfall that acts as a natural barrier for smolts and parr upstream movements. This barrier is ~3.7 km upstream of the release site, resulting in a transect of ~17.2 km. This longer transect ensured coverage of all possible upstream sites into which fish may have moved and thus reduced the possibility of undetected tags (see Table 1).

Detection efficiency

All data were initially compiled in the VEMCO VUE software, and all analyses were conducted in R version 3.5.0 (R Core Team, 2016), except if stated otherwise. For calculating detection efficiency of fixed receivers (with the exception of the last receiver downstream), fish detections at a downstream receiver were compared with those from the next upstream receiver; that is, all fish detected at receiver 2 must have passed receiver 1, providing a detection efficiency measure for receiver 1. Fixed receivers 1, 5, and 6 were included in this analysis (due to their earlier deployment, receivers 2–4 were excluded from this assessment). We also calculated the efficiency of fixed receivers by combining passive and active tracking data; this allowed for the detection of fish between receivers, improving the estimates of fixed receiver detection efficiency. Active tracking detection efficiency was

estimated by combining data from active tracking and fixed receivers. Fixed receivers were used to assess which tagged individuals remained within the study area (i.e., those that had not exited the river), and the percentage of these individuals detected by active tracking were calculated.

Sources of smolt riverine migration failure

The pattern of detection of individuals from active and passive acoustic tracking was used to infer the fate of each fish. In the process of evaluating this, we relied on two common assumptions. (1) Tag failure (e.g., through early battery discharge or mechanical failure) was assumed to be minimal. Failure of the tags used in this type of study is generally under 2% (H. H. Honkanen, unpublished data; Newton et al., 2019). (2) Tag expulsion from the fish was assumed to be inconsequential. This study accumulated data from fish over a few days to a few weeks, and where it has been examined, tag expulsion is thought to occur over a longer period (i.e., 40 days; Brunsdon et al., 2019; Lacroix et al., 2004). In addition, to infer an ultimate fate to each tagged fish, we followed postulations used in other recent studies (e.g., see Flávio et al., 2021; Gerber et al., 2017; Klinard & Matley, 2020; Villegas-Ríos et al., 2020; Weinz et al., 2020), with some modifications for the local study area, to infer nine categories of riverine migration outcomes from tag detection patterns (see Table 1 and Appendix S1: Table S1 for more details). A chi-squared test (Zar, 2010) was used to determine whether the number of migration failures was spatially randomly distributed throughout the River Endrick by dividing the river into 1-km sections.

Telemetry analyses

Because the data deviated from normality (Shapiro-Wilk test, p values <0.001), the nonparametric Wilcoxon rank tests, implemented by the `kruskalmc` function from the `pgirmess` package (Giraudoux, 2012), were used for all telemetry analyses. The four fish with fate categorized as “unknown” were excluded from further evaluation, and analyses were done separately on salmon and sea trout smolts. The duration of residency events, defined as the time that a fish spent detected at a single receiver, was compared between fish that were successful and unsuccessful river migrants and among receivers. The rate of movement, measured as the detected distance moved over time elapsed between fixed receivers, of all fish was determined, and the difference between successful and unsuccessful migrants and among different river sections

TABLE 1 The fate of each tagged fish was inferred following postulations used in other recent studies (e.g., see Flávio et al., 2021; Gerber et al., 2017; Klinard & Matley, 2020; Villegas-Ríos et al., 2020; Weinz et al., 2020), with some modifications for the local study area

Inferred riverine migration outcomes	Movement pattern criteria for inference	Species-specific
Successful	A successful river migration was assumed if tags were detected during active tracking and/or on sequential acoustic receivers downstream, to the last river receiver at the mouth of the river (Honkanen et al., 2021; Kocik et al., 2009; Lothian et al., 2018).	No
Unknown fate	Continuous upstream and downstream movements until the end of the study among tagged sea trout; the inference of migration failure (e.g., piscine predation) could not be distinguished from alternative fates (e.g., delayed migration; Thorstad et al., 2004) due to the nature of their behavior (Klinard & Matley, 2020; Villegas-Ríos et al., 2020; Weinz et al., 2020).	Yes, sea trout only
Unsuccessful river migrants: migration failure categories		
Avian predation shortly after release	If no detections were recorded by either tracking technique between the release site and the first receiver, the fish was presumed to be consumed by an avian predator. The lack of detections indicate that the tag was removed from the river system (Flávio et al., 2021; Klinard & Matley, 2020); the incidence of piscivorous mammalian predators in this river is known to be low (Malcolm MacCormick, pers. comm. 2022).	No
Avian predation	A tagged smolt was assumed to be predated by an avian predator if sequential detections from a migration downstream were followed by a sudden loss of detections by both techniques. The lack of detections indicates that the tag was removed from the river system (Flávio et al., 2021; Klinard & Matley, 2020; Villegas-Ríos et al., 2020). If no detections were recorded by both fixed acoustic receivers and active tracking for an extended period, followed by the sudden reappearance of the tag with stationary detections, recorded by either techniques, the fish was presumed to have been consumed by avian predators (i.e., removed from the river) and the tag defecated into the river (i.e., stationary signal been recorded afterward). Three tags were classified as they were defecated by avian predators, with specific detection patterns in locations or/and elapsed time to support these conclusions (see Appendix S1: Table S1).	No
Upstream mortality	Detection of a stationary tag upstream of the trap was classified as upstream mortality. In this case, the cause of the death could not be inferred; only the location of inferred mortality could be determined.	No
Tagging mortality	Early detection (i.e., within the first 7 days post-release) of a consistently stationary tag within 800 m of the release site (upstream of the fixed receiver 1) was classified as tagging mortality (Villegas-Ríos et al., 2020). Loss is inferred as direct or indirect tagging mortality from the tagging and trapping processes, the anesthetic, or/and handling.	No
Piscine predation	Inferred from an abrupt change in recorded behavior of a detected tag, often an upstream movement following the initial consistent downstream movement, we assumed an Atlantic salmon smolt was consumed by a fish predator (Flávio et al., 2021; Klinard & Matley, 2020; Villegas-Ríos et al., 2020; Weinz et al., 2020). This assumption was not made for sea trout smolts due to the nature of their behavior (e.g., Thorstad et al., 2004).	Yes, salmon only

(Continues)

TABLE 1 (Continued)

Inferred riverine migration outcomes	Movement pattern criteria for inference	Species-specific
River mortality: unknown	Where a tag was continuously detected as stationary at a location along the river, several possibilities for the fate of that fish exist: piscine defecated tag, tag expulsion, tagging mortality, or natural mortality. The mortality of the smolt was assumed (due to the stationary detection) but was classified as “river unknown” because the cause could not be inferred. The consistent tag detection implies that the tag was not removed from the river (as might be the case for avian predation), but the detection patterns were outside of the range established by our other criteria to infer a cause of loss (Flávio et al., 2021; Jepsen et al., 2019; Klinard & Matley, 2020).	No
Unknown mortality	Detections of a consistently stationary tag infer mortality, but the cause of death could not be inferred (e.g., pattern of detections outside of the range established by our other criteria, often avian with other detection patterns cannot be distinguished) (Flávio et al., 2021; Jepsen et al., 2019; Klinard & Matley, 2020).	No

Note: Riverine migration outcomes from tag detection patterns were inferred as: successful river migrant, unknown fate, and unsuccessful river migrants that were classified into seven migration failure categories. A detailed movement pattern and justification of our criteria for each fish inferred as dead or as an unknown fate is provided in Appendix S1: Table S1.

(between fixed receivers) was tested with a Kruskal-Wallis chi-squared test. To look for diurnal patterns of fish movement and any changes in these patterns, we recorded the time of day that a smolt was first detected at a receiver. Although condition factor did not vary between successful and unsuccessful smolts (Lilly et al., 2021), the effects of fork length and tag burden on smolt migration success were assessed.

RESULTS

Detection efficiency

Detection efficiency using the fixed receivers 1, 5, and 6 (Figure 1) was 98.7%, 100%, and 100%, respectively. When active tracking data were added to efficiency measures, fixed receivers 1, 5, and 6 had a detection efficiency of 95.6%, 100%, and 100%, respectively. For active tracking alone, efficiency increased with time, from 38.5% to 95.2% (Table 2). Overall, 93% (53/57) of the successful migrants, 69% (67/97) of the unsuccessful migrants, and 100% (4/4) of the unknown were detected by active tracking.

Sources of smolt riverine migration failure

In our study, 36.1% of the Atlantic salmon and sea trout smolts from the River Endrick were assumed to have made a successful river migration; that is, the tag was detected having left the River Endrick. Successful river

TABLE 2 Efficiencies of active tracking, fixed receivers, and fixed receivers encompassing active tracking data (see “Material and methods”) in the River Endrick

Date or receiver number	Percentage
Active tracking efficiency	
17 April	48
18 April	47.4
19 April	45.1
20 April	38.5
21 April	42.1
22 April	67.5
23 April	64.9
8 May	82.6
21 May	86
4 June	95.2
Fixed receiver efficiency	
1	98.7
5	100
6	100
Fixed receiver efficiency with active tracking	
1	95.6
5	100
6	100

migration for Atlantic salmon was 36.3% ($n = 49$; Figure 2) and for sea trout 34.8% ($n = 8$). 61.4% of the tagged smolts failed to migrate successfully and were

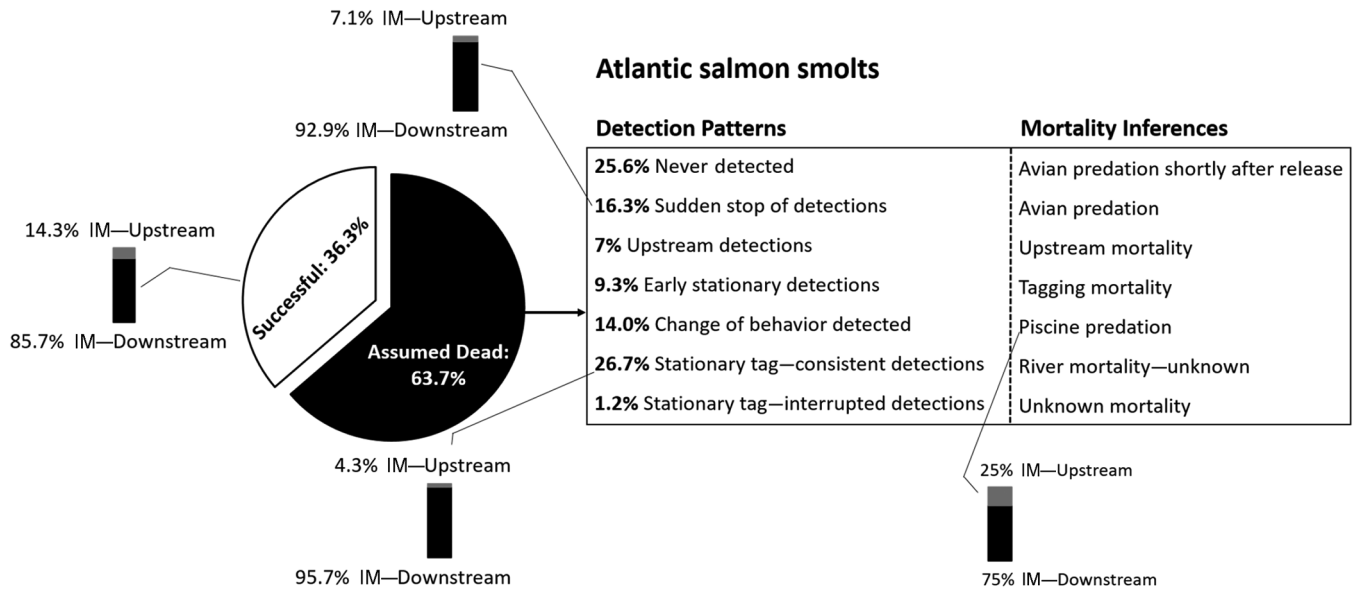


FIGURE 2 Tag movements of Atlantic salmon recorded with fixed acoustic and active tracking techniques led to the determination of detection patterns that allowed the inference of the fate for each fish (see Appendix S1: Table S1 for details). Seven detection patterns were defined to lead to a specific cause of mortality inference (see riverine migration outcomes, Table 1, and Appendix S1: Table S1). Initial upstream movement detected is reported with a bar graph. IM, initial movement

inferred as dead (see Table 1 and Appendix S1: Table S1). For Atlantic salmon, failed river migration was 63.7% ($n = 86$) and for sea trout 47.8% ($n = 11$; see Table 1 and Appendix S1: Table S1). Four sea trout smolts (17.4%) had their migration success categorized as unknown fate (see Table 1 and Appendix S1: Table S1).

Of the salmon assumed dead (sea trout excluded due to sample size; see Figure 2, Table 1, and Appendix S1: Table S1), 25.6% were never detected by both methods ($n = 22$; inferred as avian predation shortly after release, near the trap) and 16.3% of smolts suddenly stopped being detected ($n = 14$; inferred as avian predation). Thus, 41.9% of tagged salmon were thought to have been preyed upon by birds. 9.3% of the Atlantic salmon smolts ($n = 8$) were consistently detected at the same place, close to the fish release site and did not move; we infer these to be fish lost as tagging mortality. 14.0% of the Atlantic salmon smolts ($n = 12$) displayed a pattern of detection indicating a change of behavior that gave a strong indication of piscine predation. The continuous detection patterns for 26.7% of the smolts ($n = 23$) could not be classified but displayed stationary detections indicative of mortality within the river, which we categorized as river mortality—unknown. 1.2% of Atlantic salmon smolts ($n = 1$, Appendix S1: Table S1) were categorized as unknown mortality due to a stationary tag, but the detection pattern resulted in a number of possible inferences. Twenty-one percent of Atlantic salmon smolts ($n = 18$) were detected upstream of the release site. Of these, six (7%) individuals never initiated a downstream

migration, and thus, they were classified as upstream mortality. The remaining 12 individuals comprising both successful and unsuccessful migrants subsequently detected downstream of the release site (see Appendix S1: Table S1; Figure 2).

Finally, mortality-inferred events were not randomly distributed along the River Endrick ($\chi^2 = 220.54$, $df = 13$, $p < 0.001$; Appendix S1: Figure S2); the highest failure of migration was observed in the first kilometer from the release site.

Telemetry

Overall, successful river migrants of Atlantic salmon spent a median of 3.99 days (interquartile range [IQR]: 2.82–5.22) in the River Endrick, whereas sea trout spent 23.04 days (IQR: 13.98–37.75). For Atlantic salmon, the median duration of all residency events was 1156 s (IQR: 544–4109). The duration of residency events was significantly shorter for successful salmon (996 s; IQR: 500–2197) compared with unsuccessful river migrant smolts (1926 s; IQR: 635–6221; $W = 56,346$, $p < 0.01$). A similar but nonsignificant pattern was found for sea trout, with successful river migrants having shorter residency events (888.5 s; IQR: 486.5–2724.3) than unsuccessful river migrants (3431 s; IQR: 586–8842; $W = 535.5$, $p = 0.32$). The duration of residency events also varied among receivers (Kruskal-Wallis test, $\chi^2 = 29.12$, $p < 0.01$; Table 3), with smolts of Atlantic salmon displaying the

TABLE 3 Duration of residency events of smolt specific to each fixed receiver and species in the Endrick

Receiver	Species	Median duration (s)	IQR (s)
1	Salmon	1916.5	610.00–7348.00
1	Trout	2791.5	577.25–9591.25
2 ^a	Trout	179.5	128.25–230.75
3 ^a	Trout	407.5	274.25–540.75
4 ^a	Trout	744.5	502.75–1104.50
5	Salmon	804.5	441.00–2054.75
5	Trout	659.0	418.00–940.50
6	Salmon	971.0	692.00–1224.50
6	Trout	888.5	772.00–984.25
7	Salmon	1047.0	506.00–2758.50
7	Trout	1942.5	804.75–5900.50

Abbreviation: IQR, interquartile range.

^aThese receivers were deployed late and therefore only include trout movements.

longest residency around receiver 1. There were no differences in residency event durations among receivers for sea trout (Kruskal-Wallis test, $\chi^2 = 6.34$, $p = 0.096$).

For salmon, the successful river migrants had a higher rate of movement (median: 0.11 m s^{-1} ; IQR: 0.05–0.64) than unsuccessful river migrants (0.04 m s^{-1} ; IQR: 0.03–0.10; $W = 730$, $p < 0.01$). For sea trout, there was no difference in the rate of movement between successful (median: 0.25 m s^{-1} ; IQR: 0.08–0.65) and unsuccessful (0.09 m s^{-1} ; IQR: 0.04–0.36 m s^{-1} ; $W = 61$, $p = 0.53$) river migrants. For both species, the rate of movement differed among river sections (salmon: Kruskal-Wallis test, $\chi^2 = 105.95$, $df = 2$, $p < 0.01$; sea trout: Kruskal-Wallis test, $\chi^2 = 16.65$, $df = 2$, $p < 0.01$), with the movement rate increasing with distance downstream for both species (i.e., the highest movement rate being at the final section of the river, between receivers 6 and 7). For most of the migration, trout had a higher rate of movement than salmon, except for the final section (between receivers 6 and 7), whereas salmon had a higher median rate of movement than trout (0.84 and 0.71 m s^{-1} , respectively). During their migration, smolts of Atlantic salmon and sea trout passed the first receiver at all hours of the day, whereas by receivers 6 and 7, most movement was during daytime hours (Figure 3).

Although successful river migrants had a slightly larger fork length for both species (salmon: $145.3 \pm 14.5 \text{ mm}$; trout: $169.6 \pm 31.1 \text{ mm}$) than unsuccessful ones (salmon: $141.6 \pm 8.9 \text{ mm}$; trout: $168.9 \pm 36.3 \text{ mm}$), these size differences were not significant for either species (salmon: $W = 1782$, $p = 0.14$; trout: $W = 41$, $p = 0.87$). Additionally,

there were no significant differences in tag burden between successful river migrants (salmon: $5.60 \pm 1.2\%$; sea trout: $4.04 \pm 1.9\%$) and unsuccessful river migrants (salmon $5.72 \pm 1.0\%$; trout: $4.05 \pm 1.6\%$; salmon: $W = 2156$, $p = 0.82$; sea trout: $W = 44$, $p = 1$).

DISCUSSION

We provided evidence that the success of river migration by smolts in the River Endrick was low with only 36% of tagged smolts exiting the mouth of the river. This seems to be a trend emerging across rivers in Scotland (Adams, unpublished data; Honkanen et al., 2021; Lothian et al., 2018) but also elsewhere (Flávio et al., 2020; Flávio et al., 2021; Gibson et al., 2015; Jepsen et al., 2019) and other salmonid species (Welch et al., 2021; Wilson et al., 2022). Undeniably, the process of smolting is a critical phase in the life cycle of anadromous salmonids and has been associated with substantial mortality rates (Clark et al., 2016; Halfyard et al., 2012; Thorstad et al., 2012). Smolt survival can be classified as poor when less than 70% of the smolts reach the mouth of the river (Furey et al., 2016), but when riverine survival is lower than half the population, it can have drastic long-standing effects with few adults returning to spawn (Thorstad et al., 2012). Accordingly, one of the pressing requirements to better understand anadromous fish population dynamics is to be able to identify the location, the cause, and the rate of mortality that smolts are facing during their first seaward migration. Here, we produced a framework combining spatial and temporal measures of smolt riverine migration to detect and estimate mortality events using acoustic telemetry. By combining active and passive tracking techniques, we were able to refine the spatiotemporal pattern detection scale, which allowed inferences to be made about mortality location, causes, and rates that would not have been possible using fixed position acoustic receivers solely. The lack of ability to detect fish between two fixed receivers introduces uncertainties that can prevent reasonable inferences from being drawn.

Despite the importance of the smolting phase in anadromous salmonids, the migration from freshwater to marine habitats has been generally less intensively studied than adult spawning migration (Drenner et al., 2012; Furey et al., 2021). The spatial and temporal patterns of migration failure (e.g., mortality rates and location) during the smolt riverine migration have the potential to yield important insights into population bottlenecks for anadromous salmonids. In this study, we demonstrated that during their initial riverine transitional phase, smolts were particularly vulnerable to predators while migrating

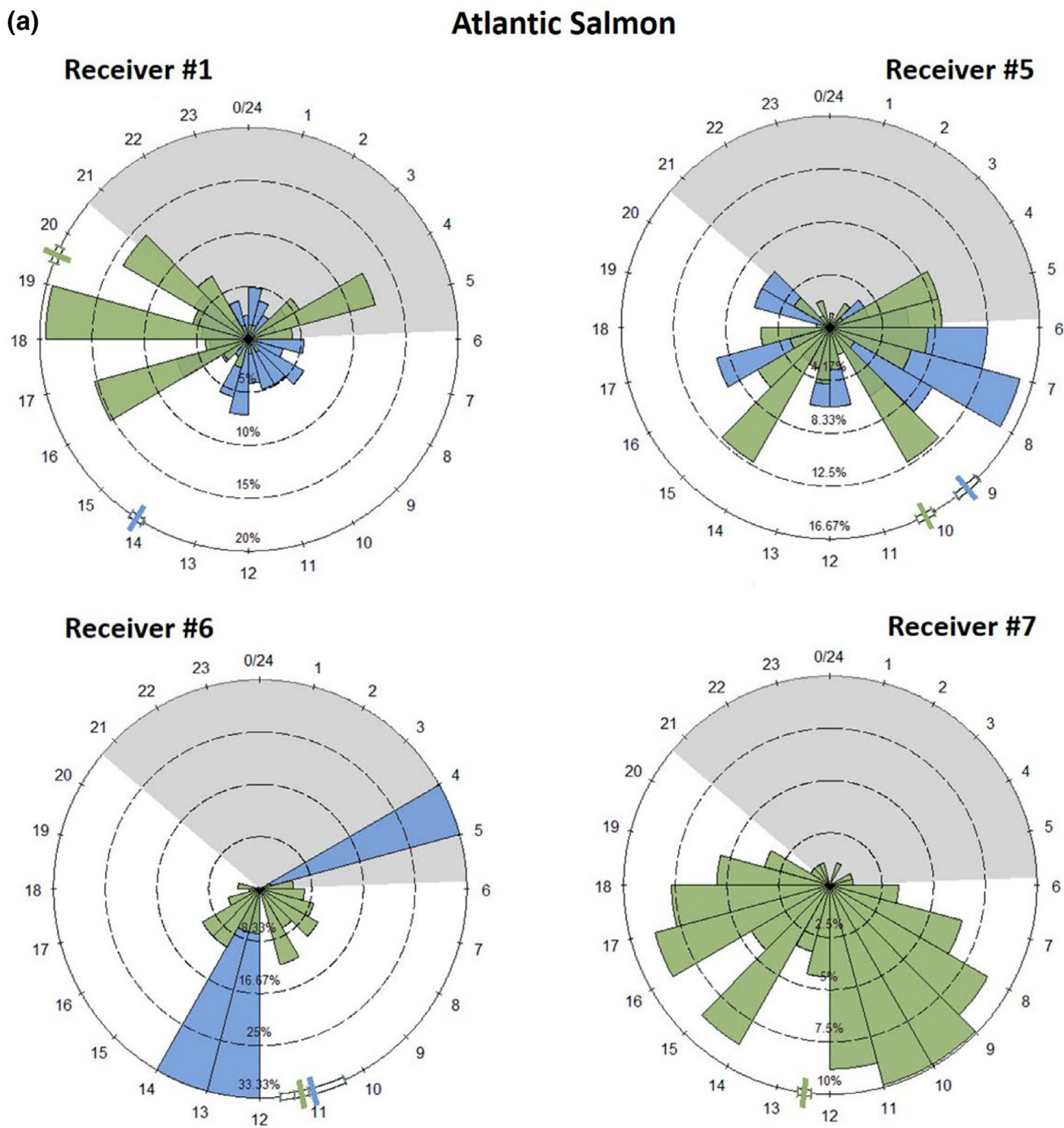


FIGURE 3 Arrival time at each fixed receiver (#1,5,6 & 7) in the River Endrick showing when Atlantic salmon (a) and sea trout (b) smolts are moving and whether the pattern changes during their migration downstream. Successful (green) and unsuccessful (blue) river migrant smolts are represented at each receiver, and each array's bar sums to 100%. The colored markers in the outer circle indicate the mean value respectively to the colored category (i.e., successful and unsuccessful river migrants). The shaded portion of the circle shows the average sunset-to-sunrise hours for April

through that high-risk landscape (Ward et al., 2008; Ward & Hvidsten, 2011). The main cause of mortality of tagged smolts in this study was interpreted as avian predation; mammalian predators are thought to be low in this study area; however, piscine predation was also

substantive. The most commonly observed avian piscine species on the River Endrick during this study were goosanders (*Mergus merganser*), gray heron (*Ardea cinerea*), and osprey (*Pandion haliaetus*), whereas the most common aquatic piscine predators in this area will likely be

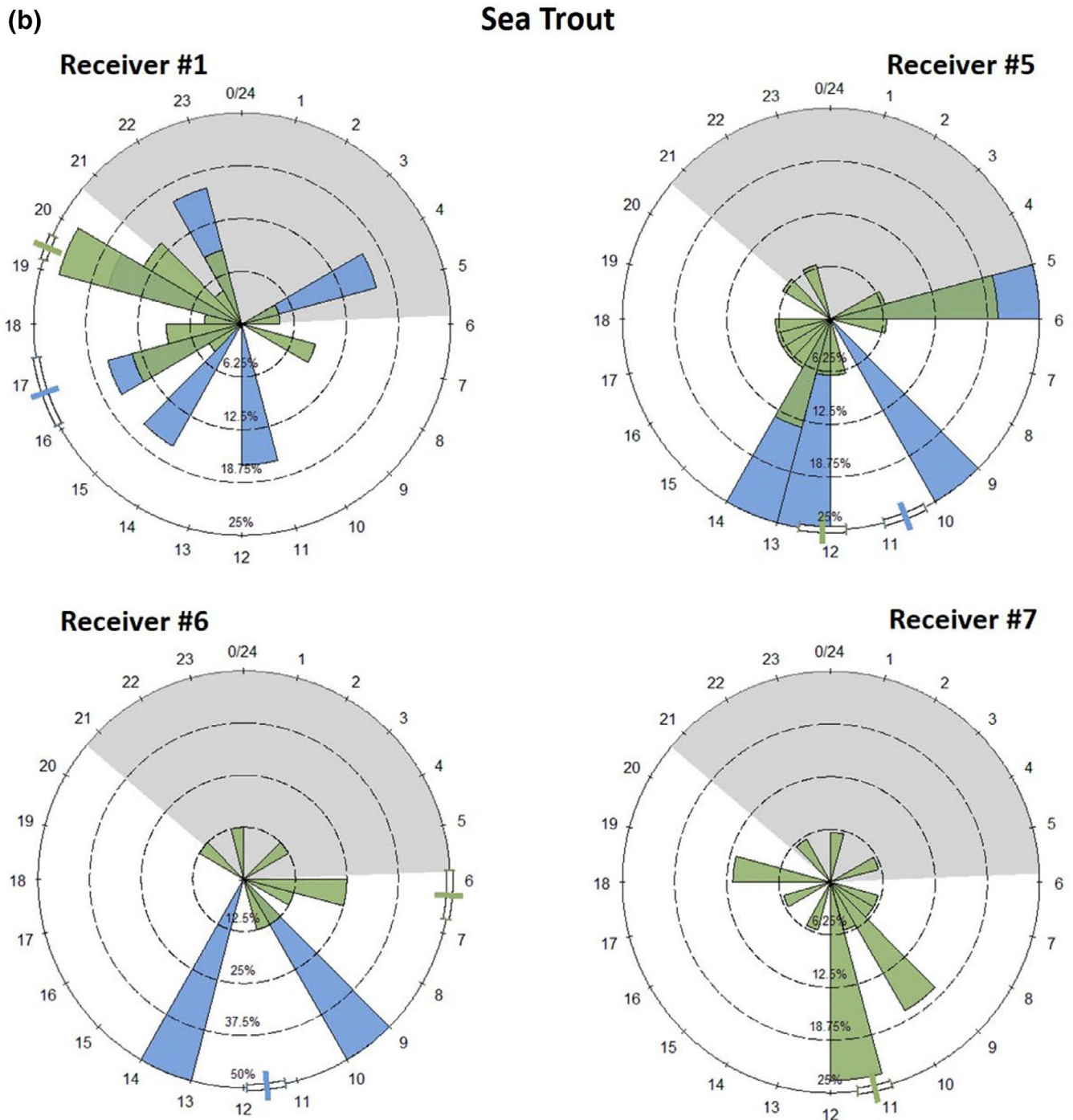


FIGURE 3 (Continued)

pike (*Esox lucius*) and brown trout (Adams, unpublished data). Predation can be a major driver of population dynamics and demography (Jonsson et al., 2017; Payton et al., 2020) by exerting a strong selective pressure (Ward & Hvidsten, 2011), but other factors can act concomitantly (e.g., environmental factors, hydropower; Lothian et al., 2018; Thorstad et al., 2012). Our finding that avian predation rates were particularly high in the

immediate vicinity of the release site suggests that there was a concomitant interaction between the effects of trapping/tagging and vulnerability to predation for smolts. Thus, the high percentage of individuals that were never detected by either method after release could be a result of either: (1) a learned response by avian predators to the numerical increase in prey near the release site (gray herons [*A. cinerea*] and goosanders

[*M. merganser*] were regularly observed close to the release site; Adams, unpublished data; Flávio et al., 2021; Ward et al., 2008; Wargo Rub & Sandford, 2020) or (2) individuals were more susceptible to predation following tagging (e.g., trapping, anesthesia, fish handling; higher residency was observed near the trap; Thedinga et al., 1994). Since the fish were released during the day, it is possible that we induced an experimental artifact by releasing vulnerable fish at a time when visually oriented predators, such as birds, are particularly effective. In a previous study in a different Scottish river, there was no difference in survival of smolt release during the day versus night (C. E. Adams, unpublished data). Nevertheless, other explanations independent of trapping/tagging-induced effects, such as a naturally high predation level in that specific area, cannot be ruled out.

A contentious point of using telemetry has always been to estimate the direct effect of tagging on smolt fate, which has been mostly tested experimentally or using models, but rarely in the wild (Hueter et al., 2006; Klinard & Matley, 2020; Newton et al., 2016; Vollset et al., 2020). This study suggests some tagging-induced mortality, but it was unrelated to tag burden. The varying skills of the tagger when inserting tags surgically have been shown to have an effect on the outcome of the tagging process (Cooke et al., 2003), but environmental factors, such as elevated water temperature, can also have an effect on tagging outcomes (Brownscombe et al., 2019). One way or the other, this result highlights the urgent need to determine direct and indirect effects of surgically implanted tags in telemetry studies to ensure adequate interpretation of the results.

Tagged smolts showed migratory behavioral patterns that were both similar and distinct to other migratory salmonid populations experiencing high predation risks. Residence period and rate of movement of smolts while migrating indicated that a shorter time spent within the river was related to migration success, corroborating the notion that the River Endrick is a high-risk landscape for salmonid juveniles (Furey et al., 2016; Furey et al., 2021; Honkanen et al., 2018). In this study, we did not directly examine the physiological status of migrating smolts. Condition factor might arguably provide an indication of gross physiological condition. However, in this study there was no difference in condition factor between successful and unsuccessful river migrants (Lilly et al., 2021), which suggest that the process of smolting was well developed in all individuals used in the study and that overall body condition was not a determinant of migration success. We cannot, however, rule out other physiological drivers of river migration success from this study (McCormick et al., 2009; Thorstad et al., 2012).

Despite the evidence of high predation pressure found in the river, smolts traveled during the day, especially as

they moved further downstream. Typically, smolts travel during the night as a behavioral predation risk-reduction mechanism (Flávio et al., 2020; Furey et al., 2016; Ibbotson et al., 2006). Diurnal traveling is counterintuitive to predation avoidance, particularly in a river with visual predators, and thus, migration during the night should lead to lower mortality. However, other studies have also found a diurnal pattern (Fängstam et al., 1993; Thorstad et al., 2012), indicating the occurrence of daytime migration could be related to other factors (e.g., the predator community of a river, latitude, and how early or late in the period of the smolt migration period the fish is).

Upstream movements are considered unlikely for migrating Atlantic salmon smolts, and downstream movements are assumed to be ubiquitous (Gauld et al., 2013; Lothian et al., 2018). Yet, movements of the smolts in the River Endrick offered a diverse range of upstream movements, with a variation in time and final direction (downstream or upstream). Salmon telemetry studies rarely link upstream movements to tag effects or handling and to mortality events occurring upstream after tagging (Frank et al., 2009). These complex behaviors could also be induced from exploratory movements (Keefer et al., 2008), seeking alternative routes, waiting for appropriate conditions (Holbrook et al., 2009), disorientation in certain hydraulic conditions (Honkanen et al., 2021), or varying sensitivity in distinct migratory phases (Mäkinen et al., 2000). The amount of upstream movement and mortality found in this study highlight the importance of monitoring these movements to ensure adequate interpretation of the results of telemetry studies.

Important considerations

A few caveats should be noted that could alter interpretations in this study. Despite our best efforts, some uncertainties persisted when mortality sources and rates were untangled and inferred. For example, a number of uncertainties are related to the detection patterns of “never been detected” or “sudden stop in detections,” which were inferred as avian predation. Other factors could induce these patterns, that is, undetected long-distance upstream movements (no fixed receivers were positioned upstream), mammal predation, and/or tag malfunctioning/expulsion that could result in an overestimation of the avian predation. Although bias in the precision in estimating the proportion of fish in any inferred mortality category is probable, regardless of the uncertainties in our estimations, the main point remaining here is that by far the most substantive detection pattern points to a high percentage of avian predation events.

Our result bias can also be related to underestimating mortality events, which is likely to occur in the delineation of piscine predation events. Due to the difficulty in characterizing a change of behavior, we included only the most obvious patterns from our detections. Most likely, a number of piscine predation mortality events were classified as river mortality—unknown—because of the difficulty in distinguishing predator from prey behaviors (vice versa is also possible, a prey behavior could have been quantified as a predator behavior, although this is less likely due to the strong unidirectional downstream migration patterns displayed by Atlantic salmon). There are also uncertainties associated with the location and timing of these piscine predation events (i.e., where and when the smolt was predated). The use of predation tags (Weinz et al., 2020), which can detect predation events, coupled with active tracking, could be a valuable method to further refine mortality events of smolts. Finally, due to our strict temporal and spatial assumptions around mortality near the release site, the inferred tagging mortality rate might be an underestimation. In reality, mortality events associated with tagging could take place downstream and upstream of the release site and/or after 7 days; in this paper, these events have been classified as river mortality or upstream mortality.

There was a learning curve associated with the active tracking methodology, mostly linked to the disruptive code collisions (i.e., detection interference as a result of several transmissions echoes being heard simultaneously; Binder et al., 2016). Because 53 of 57 successful migrants were detected by active tracking, the increase in detection efficiency over time appears to be linked to the gained experience of handling tag transmission collisions than differences in detection between alive and dead individuals. Even with the poor performance of active tracking at the start of the study, detection efficiency of active tracking improved and refined the scale of detection when those data were combined with fixed receivers.

Despite these caveats, the aim of this paper was to disentangle some aspects of the location, cause, and rate of mortalities occurring and influencing riverine migration success based on several published assumptions with their associated uncertainties (enumerated above). These assumptions are frequently used by studies to target specific mortality causes independently (e.g., avian and piscine predation). Combining spatial and temporal metrics allowed us to attempt to improve the resolution of the tagged smolt fates, while we are acknowledging some uncertainties surrounding identifying location, cause, and rate of mortality events remain. As more studies refine active tracking combined with fixed receivers as a tracking methodology in the riverine environment, assigning location, cause, and rate of mortality more precisely is anticipated.

CONCLUSIONS

Active tracking combined with fixed acoustic receivers provided considerable insights into the patterns of migration failure and their causes that could not be identified by relying solely on fixed acoustic receivers. Refining the temporal and spatial scale of the mortality location enabled the causes and rates to be inferred, which we embedded this into a behavior and movement study. Thus, we provided a case study of a methodology, applicable to any small-moderate riverine system, which should stimulate further use in an emerging research theme and provide insights to help to guide management actions (Flávio et al., 2021; Klinard & Matley, 2020; Villegas-Ríos et al., 2020). The extended spatial and temporal characteristics of salmonid migration mean that factors acting over long periods and broad geographic scales may all contribute, both cumulatively and synergistically, to the currently depressed populations (Knudsen & McDonald, 2020; Thorstad et al., 2012). Thus, the freshwater and marine survival of anadromous salmonids are known to be intertwined at the population level, although our knowledge of the relationship between the two is limited (McCormick et al., 2009; Sobocinski et al., 2021). Anadromous salmonid marine mortality is considered to be density-independent (Flávio et al., 2019; Jonsson et al., 1998), which implies that the number of juvenile individuals, the smolts, entering seawater will broadly determine the number of adults that return from sea to spawn (McCormick et al., 2009). Thus, success in the early stages of migration to sea success is likely to have far-reaching effects on population dynamics. Considering the major declines of anadromous salmonid worldwide (Chaput, 2012), anadromous salmonid populations are of high conservation and scientific relevance, and identifying bottlenecks during their freshwater transition to the marine environment should allow specific management actions to mitigate these losses.

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CONFLICT OF INTEREST

The authors declare no conflict of interest.

AUTHOR CONTRIBUTIONS

Louise Chavarie conceived the study. Colin E. Adams funded the study. Louise Chavarie, Hannele M. Honkanen, Jessie M. Lilly, Hannah R. Greetham, and Colin E. Adams carried out the fieldwork. Louise Chavarie, Hannele M. Honkanen, Jessie M. Lilly, and Matthew Newton participated in the data analyses. Louise Chavarie wrote the manuscript. All authors read, edited and approved the final manuscript.

DATA AVAILABILITY STATEMENT

Data and code (Chavarie, 2022a, 2022b), respectively, are available from Figshare: <https://doi.org/10.6084/m9.figshare.19153859.v1> and <https://doi.org/10.6084/m9.figshare.19127534>.

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SUPPORTING INFORMATION

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